

DOI: 10.17951/pjss/2017.50.2.237

ADAM BOGACZ*, BARTŁOMIEJ GLINA**, KATARZYNA
JANUSZEWSKA*, ANNA MIDOR*, KLARA TOMASZEWSKA***,
MAGDA PODLASKA***, PAWEŁ TELEGA*,
AGNIESZKA DRADRACH****

THE EFFECT OF LONG-TERM PEATLANDS DRAINAGE
ON THE PROPERTIES OF SOILS IN MICRORELIEF IN THE
DŁUGIE MOKRADŁO BOG (CENTRAL SUDETES – SW POLAND)

Received: 13.07.2017

Accepted: 30.11.2017

Abstract. The aim of the study was to assess the impact of long-term drainage on the morphology and selected properties of shallow peatland soils in microrelief. The study was conducted within a strongly drained peatland (Długie Mokradło bog) located on an elevated plateau in the Central Sudetes. The study area is covered by spruce stands introduced by man. Long-term drainage has changed the morphology of study soils which were classified as Ombric Fibric Dystric Histosols or Histic Dystric Gleysols. Some peat horizons were strongly silted. The depth of organic materials varied within the range of 30–55 cm. Peat humification process showed greater activity in surface horizons than in deeper ones. This phenomenon was especially visible in the shallow places in drainage ditches. Soil reaction was strongly acidic. In soil horizons, in old drainage ditches, high-

* Wrocław University of Environmental and Life Sciences, Institute of Soil Science and Environmental Protection, Grunwaldzka 53, 50-357 Wrocław, corresponding author: adam.bogacz@edu.gmail.com.

** Poznań University of Life Sciences, Department of Soil Science and Land Protection, Szydlowska 50, 60-656 Poznań.

*** Wrocław University of Environmental and Life Sciences, Institute of Botany and Plant Ecology, Grunwaldzka 24a, 53-363 Wrocław.

**** Wrocław University of Environmental and Life Sciences, Institute of Agroecosystems and Green Areas, Grunwaldzka 24a, 53-363 Wrocław.

er values of effective cation exchange capacity (CECe) were recorded, whereas base saturation (BS) did not exceed 20%.

Keywords: peat soil, mountain peatland, Sudetes, old ditches, peat cores morphology

INTRODUCTION

Drying of peatlands located on many plateau areas in the Stołowe Mountains at the turn of the 19th and 20th centuries led to the changes of landscape morphology in the macro- as well as in the microrelief (Migoń *et al.* 2011). Drained peatland is a labile ecosystem closely linked to groundwater levels and to changing vegetation structure (Hokka *et al.* 2008, Sarkkola *et al.* 2010). Processes occurring within drainage peatlands such as subsidence, reducing the depth of peat, its mineralization were affected by condition and functioning of drainage ditches (Laiho 2008). Among ditch trenching factors there are also: organic matter slime (Sallantaus 1988, Prevost *et al.* 1999) and periodic bank ditch washing by retaining water (Rantonen and Paivanen 1999) – in the steep area the collapse of ditch walls is often observed, furthermore it might be also the consequences of peat freezing and erosion (Berry and Jeglum 1991). Water takes part of the soil from the walls and bottom of the ditches. The shallow organic soils are especially exposed to erosive drainage. Because of vegetation, shallow drainage ditches are exposed to quick overgrowing (Minayeva *et al.* 2009, Marttila and Klove 2010). The blocking of water flow in channels by sediments, and various groups of plants (e.g. mosses, sedges, grasses) intensified this process (Varry 1988). Soil material accumulated in lower ground forms is usually enriched in macronutrients in comparison to material occurring between ditches (Pietilainen and Rekolainen 1991, Joensuu *et al.* 1999).

The aim of the work was to assess the current structure and properties of soils within the strongly drained area, after more than 100 years of implementation of land melioration. In this study the comparison of soils in drainage ditches and the area between them (mounds) were done.

MATERIALS AND METHODS

The study was carried out on the Długie Mokradło bog located in the Skalniak plateau. Study peatland is covered by *Calamagrostio villosae-Piceetum* (L.) and sphagnum subgroup *Calamagrostio villosae-Piceetum sphagnetosum* (L.) class (Glina 2014). Within the study area two sampling sites were established. Sampling site 1 (16°17'28.1"E, 50°28'28.7"N) included a fragment of a 20-year-old spruce monoculture, whereas sampling site 2 (16°17'43.8"E, 50°28'25.1"N) – the fragment of 70–90-year-old spruce stands. The peatland

surfaces were “cut” with a shallow drainage system of 2 m × 2 m spacing. This type of drainage system was rarely used in the past in peatlands without trees in Norway (Braekke 1978). The drainage was generally made to a depth of 50 cm or – in some cases – to mineral bedrock. The total length of ditches in the Stołowe Mountains National Park is ca. 250 km (Ciężkowski and Kielczawa 2008). The drainage ditches system and the location of the peat cores within the study area is visible in the LIDAR photo (Fig.1) (geoportal.gov.pl).



Fig. 1. Location of Stołowe Mountains a – Długie Mokradło bog, b – location of sampling sites; 1, 2 – number of fields

Peatland under study is supplied with water primarily by atmospheric deposition (ombrogenic type) and partially (the edge parts) by local stream Czermnica (fluviogenic type) (Jermaczek *et al.* 2012). In each sampling sites, five peat cores were retrieved and described. Half of them represented soils from old drainage ditches, nowadays overgrown by plants and filled with peat. The rest of cores were collected from the mounds between ditches (Fig. 2). Soil material for laboratory analysis was sampled by genetic soil horizons, packed into the plastic bags and transported to laboratory. In total, 48 soil samples (40 organic and 8 mineral) were collected. In fresh (moist) materials, following properties were determined: peat decomposition rate using SPEC method (Lynn *et al.* 1974), degree of secondary peat transformation by coefficient W_1 index by Gawlik (2000), pH in water and 1 mol·dcm⁻³ KCl (ratio 1:2.5). Whereas in the dry material ash content after placing dried samples for 5h in a muffle furnace (Bojko and Kabala 2014), the hydrophobic potential using MED method (Doerr 1998), content of acid cations

(H^+ , Al^{3+}) extracted with $1\text{ mol}\cdot\text{dm}^{-3}$ HCl using titration methods, the content of exchangeable base cations (S) (Ca^{+2} , Mg^{+2} , K^+ , Na^+) extracted with $1\text{ mol}\cdot\text{dm}^{-3}$ CH_3COONH_4 at pH 7.0 using AAS analyzer (Van Reeuwijk 2002) were determined. The effective cation exchange capacity (CECe) and base saturation (BS) were calculated. The more detailed characterization of the peat soils from the Długie Mokradło bog was reported by Głina *et al.* (2016, 2017).

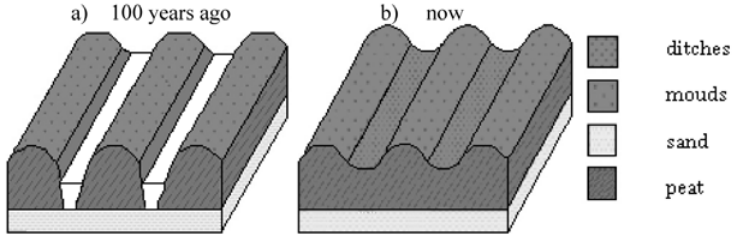


Fig. 2. Microrelief of the sampling sites

RESULTS AND DISCUSSION

Morphology of soil profile

Study soil consisted of hemic and fibric material in particular. The layers of sapric peat material were observed directly over sandy bedrock or near the soil surface in a sampling site 1 (Fig. 3 and Fig. 4). Thickness of organic materials did not exceed 50 cm. Soils formed the ditches which were often less different from the neighboring soils, showing generally moderate to low degree of peat decomposition. Soil materials of higher degree of decomposition were accumulated generally in ditches. It is the result of moisture conditions changes caused by surface run off (Chambers 1983). In deeper soil horizons, higher contribution of sand material was recorded than in upper horizons (Table 1). This situation might be the result of drainage ditches installation, which bottom reached the mineral materials (Joensuu *et al.* 2002).

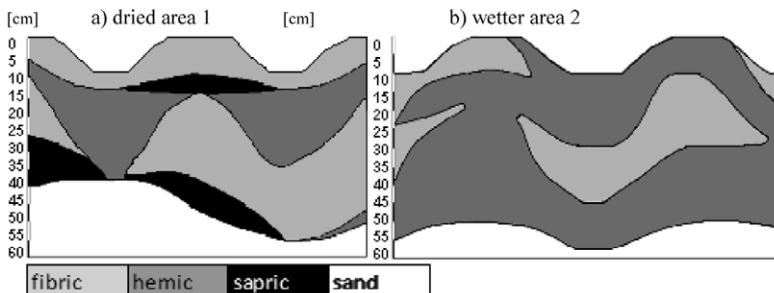


Fig. 3 and 4. Morphology of organic soils profiles in old drainage areas (current state)

Sand admixture in the surface horizons can also be associated with deforestation of adjacent areas, leading to increased erosion (Jermaczek *et al.* 2012). Plowing shallow ditches and furrow deposition also contributed to the movement of soil material, creating numerous horizons of low density (Table 1, Fig. 1 and Fig. 4). Earlier studies conducted by Bogacz and Roszkowicz (2010) in the Kragłe Mokradło bog showed the occurrence of thick (several centimeters) sandy layers in the organic soils. It was most probably related to the changes in the morphology of the drained areas (Ingram 1992) or the different speed of transportation and deposition of mineral and organic materials in ditches (Van Rijn 1987). The plant material that builds the organic levels of the soils of two research surfaces 1 and 2 were mainly the remains of plants of the genera: *Sphagnum* (L.) and *Eriophorum* (L.). Sometimes these materials also included remains of wood and bark fragments. The remains of *Polytrichum sp.* (L.) were also found (Glina *et al.* 2017).

Physical properties of soil

More than half of the organic samples in the drainage areas are not silted according to Okruszko (1993), containing less than 25% of the mineral material in dry mass of sample. Low ash content is determined by the dominance of moss vegetation (Pakarinen and Gorham 1984) and type of water supply of soil, mainly by precipitation. On the analyzed areas, there was observed tendency to increase the ash content with depth in the profile ($r=0.45^*$, $n=48$, $p<0.05$) (Table 3). The highest values of particle density above $2.00 \text{ g}\cdot\text{cm}^{-3}$ were calculated for peat with sand materials admixture and strong degree of organic matter decomposition (Table 1). This data are converging with the research provided by Wang *et al.* (2016) on mountain organic soils with both underlying and interbedded mineral material. However, recorded values of particle density slightly exceeded $1.50 \text{ g}\cdot\text{cm}^{-3}$. Bulk density is strongly associated with ash content, peat botanical composition degree of peat decomposition and drying (Nichols and Boelter 1984). In study soils, bulk density varied within a narrow range from 0.10 to $0.40 \text{ g}\cdot\text{cm}^{-3}$ (Table 1). The total porosity (P_c) of organic horizon varied widely in the profile and ranged from 82.5 to 93.1% of soil volume. The molarities ethanol droplet test (MED) revealed strong to extreme hydrophobicity of organic horizons. The litter horizons (O_l) and slightly decomposed peat (O_i) were generally stronger water repellent than others. More hydrophobic properties represented organic horizons in profiles (1–5) from mounds rather than (6–10) from ditches (Table 1). Research conducted by Łachacz *et al.* (2009) on numerous samples of the post boggy soils in the north-eastern part of Poland indicated the dependences between the values of the MED test, type of peat, content of organic matter and, above all, the stages of its decomposition. There were different stages of secondary transformation (index W_1) of organic matter in the studied drained area (Table 1). The values of W_1 index ranged from 0.26

TABLE 1. PHYSICAL PROPERTIES OF SOIL

N°	Soil horizon	Depth (cm)	Ash content (%)	Profile				RF	PI	MED (%)	W ₁
				ρ_w (g·cm ⁻³)	ρ_o (%)	Pc (%)	(%)				
mounds	O1	4-9 6	4.81-8.67 6.34	1.50-1.55 1.52	0.11-0.13 0.12	91.9-92.7 92.4	50-66 57	5 5.0	24-36 32	0.66-0.71 0.68	
	Oi	7-26 13	3.57-32.3 10.2	1.50-1.80 1.55	0.11-0.22 0.13	82.5-93.1 91.8	52-78 63	3-6 5.0	24-36 26	0.26-0.59 0.41	
	Oe	4-15 9	2.51-48.4 19.2	1.50-1.98 1.66	0.10-0.28 0.16	82.5-93.1 89.3	18-40 29	0-6 3.4	24-36 25	0.39-0.55 0.44	
ditches	Oa	5-15 10	22.4-77.8 52.6	1.70-2.31 2.03	0.18-0.40 0.30	82.6-89.3 85.5	7-16 12	1-3 2.0	24-26 25	0.36-0.61 0.46	
	O1	3-4 3	6.76-10.9 8.83	1.50-1.53 1.52	0.11-0.12 0.11	92.2-92.7 92.4	53-64 59	5-6 5.5	24-36 30	0.67-0.79 0.73	
	Oi	4-23 14	4.68-23.4 11.8	1.50-1.70 1.59	0.11-0.18 0.14	89.4-92.7 91.3	44-87 55	5-7 6.0	13-36 24	0.41-0.89 0.56	
6-10	Oe	5-17 10	7.4-69.9 24.7	1.53-2.19 1.72	0.12-0.37 0.20	83.3-92.3 89.5	26-40 35	2-6 4.1	24-36 27	0.31-0.69 0.47	

TABLE 2. PHYSICO-CHEMICAL PROPERTIES OF SOIL

N°	Soil horizon	pH H ₂ O	Profile							CECe	BS (%)
			H ⁺ +Al ³⁺	Ca ²⁺	Mg ²⁺	K ⁺ (cmol(+) kg ⁻¹)		Na ⁺	S		
mounds	OI	<u>3.9-4.4</u> 4.1	<u>12.8-16.0</u> 14.4	<u>0.96-1.20</u> 1.09	<u>1.14-1.27</u> 1.20	<u>0.37-0.80</u> 0.59	<u>0.16-0.20</u> 0.18	<u>2.90-3.30</u> 3.07	<u>15.7-19.3</u> 17.5	<u>17.0-18.5</u> 17.6	
	Oi	<u>3.2-4.0</u> 3.7	<u>23.2-29.6</u> 25.6	<u>0.87-2.0</u> 1.41	<u>0.25-1.16</u> 0.69	<u>0.09-0.40</u> 0.16	<u>0.06-0.25</u> 0.17	<u>1.60-3.40</u> 2.43	<u>24.8-31.8</u> 28.1	<u>6.4-13.0</u> 8.7	
	Oe	<u>3.3-4.5</u> 3.8	<u>11.6-28.0</u> 22.6	<u>0.95-5.74</u> 2.50	<u>0.10-1.41</u> 0.79	<u>0.09-0.50</u> 0.23	<u>0.04-0.22</u> 0.15	<u>1.59-7.70</u> 6.93	<u>13.3-30.8</u> 23.6	<u>6.6-26.2</u> 14.6	
1-5	Oa	<u>3.3-3.4</u> 3.3	<u>8.0-23.2</u> 17.1	<u>0.80-1.20</u> 0.99	<u>0.52-0.69</u> 0.61	<u>0.07-0.22</u> 0.12	<u>0.12-0.17</u> 0.15	<u>1.70-2.10</u> 1.90	<u>9.70-25.1</u> 19.0	<u>7.4-17.3</u> 11.4	
	OI	<u>4.5-4.7</u> 4.6	<u>12.0-24.0</u> 18.0	<u>1.20-1.60</u> 1.40	<u>0.91-1.24</u> 1.08	<u>0.60-0.70</u> 0.65	<u>0.20</u> 0.20	<u>2.90-3.70</u> 3.30	<u>15.7-26.9</u> 21.3	<u>10.8-23.8</u> 17.3	
ditches	Oi	<u>3.6-4.4</u> 4.1	<u>24.8-32.2</u> 29.8	<u>0.79-3.99</u> 2.15	<u>0.16-1.13</u> 0.50	<u>0.05-0.45</u> 0.32	<u>0.14-0.65</u> 0.30	<u>1.44-5.40</u> 3.28	<u>30.2-35.8</u> 33.0	<u>6.5-17.9</u> 10.6	
	Oe	<u>3.1-4.5</u> 3.8	<u>13.6-40.0</u> 27.4	<u>0.95-3.83</u> 1.69	<u>0.16-1.06</u> 0.50	<u>0.07-0.35</u> 0.19	<u>0.06-0.21</u> 0.16	<u>1.32-4.50</u> 2.54	<u>15.8-38.8</u> 28.7	<u>6.8-14.1</u> 8.9	

in fibric peat horizons to 0.89 in the fibric material enriched with sand. The initially and weakly transformed peat is dominant on the elevated areas (mounds), while former drainage ditches were filled with strongly secondary transformed peat (Fig. 4). W_1 index indicated a much lower degree of transformation of the soil from drained forestry area of the Sudetes than the organic soils of other drainage areas in the 1950–1960s (Matyka-Sarzyńska and Sokołowska 2005, Kalisz *et al.* 2015). The above statements were also confirmed by Glina *et al.* (2016).

The strongly acidic reaction of all soil horizons, as well as their high values of exchangeable acidity ($H^+ + Al^{+3}$), were favored high effective exchange cation capacity (CECe) and low values of base saturation (BS) ($r=0.41^*$, $n=40$, $p<0.05$). In wet parts such as ditches, slightly higher pH values were observed rather than in soils from mounds. Drainage of peat soils intensified the process of soil acidification especially in areas between the ditches (Minkinen and Laine 1996). Peat aeration changed the microbe activation and degree of peat decomposition (Clymo 1983). Low value of soil pH was associated with a lower degree of organic matter hummification, expressed by the index (PI) ($r=0.53^*$, $n=40$, $p<0.05$). There was an increase of the trophy status of soils in ditches, when compared to the soils formed on the mound. Mean values of the (CECe) and (S) were slightly higher in soil from the ditches than in the mounds (Table 2). This is probably due to the concentration of solid materials in the ditches (Joensuu *et al.* 1999) and, additionally, the effect of mineral bedrock in case of shallow organic horizons (Nieminen *et al.* 2010). Values of CECe were negatively correlated with the degree of peat transformation and ash content (Table 3). The base saturation (BS) is low in all tested soils and the value of this factor for individual horizons in general did not exceed 20%. Proportion of base cations (S) was the highest in litter and other more hydrophobic (MED) ($r=0.38^*$, $n=40$, $p<0.05$) and stronger transformed W_1 ($r=0.43^*$, $n=40$, $p<0.05$) (Table 3) horizons.

TABLE 3. CORRELATION COEFFICIENT BETWEEN SELECTED PROPERTIES OF ORGANIC HORIZONS

Value	pH	W_1	PI	MED	Ash	RF	CECe	BS
Depth	-0.06	-0.19	0.12	0.25	0.45*	-0.09	-0.13	-0.02
pH		0.38*	0.53*	0.12	-0.34*	0.23	-0.06	0.41*
W_1			0.17	-0.02	0.11	0.18	-0.34*	0.43*
PI				0.25	-0.37*	0.64*	0.05	0.25
MED					-0.17	0.33	-0.18	0.38*
Ash						-0.53*	-0.55*	0.00
RF							0.04	0.06
CECe								-0.44*

Correlation ratio at: * $p<0.05$, $p<0.01$, $n=40$

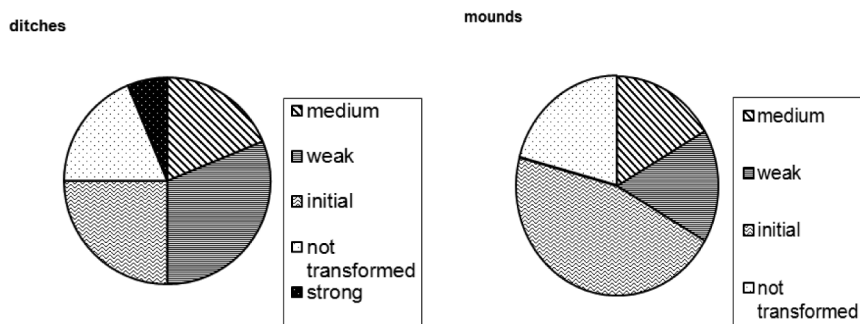


Fig. 5. State of secondary transformation of the organic soils horizons
a) ditches and b) mounds

CONCLUSIONS

1. The old drainage system has changed the soil morphology and increased the diversity of soil horizons, especially on a drier area.
2. Peatland forestry has transformed organic horizons more in drainage ditches than in parts called mounds.
3. The development of several centimeters of litter *folic* horizons may indicate the inhibition of peat forming process in the dry part of peatland.
4. Shallow Histosols and Histic Gleysols showed a greater trophy status in soil horizons from old drainage ditches than in the mounds.

REFERENCES

- [1] Berry, G.J., Jeglum, J.K., 1991. *Hydrology of drained and undrained black spruce peatlands: ground water table profiles and fluctuation*. Forestry Canada, Ontario Region, COFRA Report 3307: 1–31.
- [2] Bogacz, A., Roszkowicz, M., 2010. *Influence of forest management on the changes of organic soil properties in marginal part of Kragle Mokradlo Peatlands (Stolowe Mountains National Park)* (in Polish). *Soil Science Annual*, 61, 2: 15–20.
- [3] Bojko, O., Kabala, C., 2014. *Loss-on-ignition as an estimate of total organic carbon in the mountain soils*. *Polish Journal of Soil Science*, 47, 2: 71–79.
- [4] Braekke, F.H., 1978. *Afforestation on peatland in Norway. Proceedings of the IPS-Symposium, International Peat Society, Brumunddal, Norway*, pp. 73–96.
- [5] Chambers, F., 1983. *New application paleo ecological techniques: integrating evidence of arable activity in pollen peat and soil stratigraphy, Cefn Graenog, North Wales*. [In:] M. Jones (ed.), *Integrating the Subsistence Economy*, BAR International Series 181, Chapter 5, 16 pp.
- [6] Ciężkowski, W., Kielczawa, B., 2008. *Wody* (in Polish). [In:] A. Witkowski, B.M. Pokryszko, W. Ciężkowski (eds.), *Przyroda Parku Narodowego Gór Stołowych*. Wyd. Parku Narodowego Gór Stołowych, Kudowa-Zdrój, pp. 86–100.
- [7] Clymo, R.S., 1983. *Peat. Ecosystems of the World: Bog, Swamp, Moor and Fen*, vol. 4A, A.J.P. Gore (ed.), Amsterdam, Elsevier, pp. 159–224.

- [8] Doerr, S.H., 1998. *On standardizing the Water Drop Penetrator Time and the Molarity of an Ethanol Droplet technique classify soil hydrophobicity: a case study using medium texture soils*. *Earth Surface Processes Landforms*, 23: 663–668, DOI: 0.1002/(SICI)1096-9837(199807)23:7<663::AID-ESP909>3.0.CO; 2–6.
- [9] Gawlik, J., 2000. *Division of differently sited peat formation into classes according to their state of secondary transformations* (in Polish). *Acta Agrophysica*, 26: 17–24.
- [10] Glina, B., 2014. *Spatial diversity of shallow organic soils in the Stołowe Mountains as a result of anthropogenic transformation*. Wrocław University of Life Sciences, Ph.D thesis, Wrocław, Poland. 208 pp.
- [11] Glina, B., Bogacz, A., Gulyas, M., Zawieja, B., Gajewski, P., Kaczmarek, Z., 2016. *Effect of long term forestry drainage on the current state of peatland soils. A case study from the Central Sudetes, SW Poland*. *Mires and Peat*, 18, 21: 1–11, DOI: 10.19189/Map.2016.OMB.240.
- [12] Glina, B., Malkiewicz, M., Mendyk, Ł., Bogacz, A., Woźniczka, P., 2017. *Human affected disturbances in vegetation cover and peatland development in the Holocene recorded in shallow mountain peatlands (Central Sudetes, SW Poland)*. *Boreas*, 46, 2: 381–390, DOI: 10.1111/bor.12203.
- [13] Hokka, H., Repola, J., Laine J., 2008. *Quantifying the interrelationship between tree stand growth rate and water table level in drained peatlands sites within Central Finland*. *Canadian Journal of Forest Research* 38, 7: 1775–1783, DOI: 10.1139/X08-028.
- [14] Ingram, H.A.P., 1992. *Introduction to the ecohydrology of mires in the context of cultural perturbation*. [In:] O.M. Brag, P.D. Hulme, H.A.P. Ingram, R.A. Robertson (eds.), *Peatlands Ecosystem and Man: an Impact Assessment*, International Peat Society Finland, 676–693.
- [15] Jermaczek, A., Wolejko, L., Chapiński, P., 2012. *Mokradła Sudetów Środkowych i ich ochrona*. Wydawnictwo Klubu Przyrodników, Świebodzin 51 pp. (in Polish).
- [16] Joensuu, S., Ahti, E., Vuollekoski, M., 2002. *Effects of Ditch Network Maintenance on the Chemistry of Run-off Water from Peatland Forests*. *Scandinavian Journal of Forest Research*, 17, 3: 238–247, DOI: 10.1080/028275802753742909.
- [17] Joensuu, S., Ahti, E., Vuollekoski, M., 1999. *The effects of peatland forest ditch maintenance on suspended solids in run off*. *Boreal Environmental Research* 4: 335–343.
- [18] Kalisz, B., Łachacz, A., Głazewski, R., 2015. *Effects of peat drainage on labile organic carbon and water repellency in NE Poland*. *Turkish Journal of Agriculture and Forestry*, 39: 20–27.
- [19] Laiho, R., 2008. *From a mire ecosystem to a drained peatlands*. [In:] R. Korhonen, L. Korpela, S. Sarkkola (eds.), *Finland-Fenland. Research and sustainable utilization of mires and peat*. Finnish Peatland Society and Maahenki Ltd. 152–157.
- [20] Lynn, W.C., McKinzie, W.E., Grossman, R.B., 1974. *Field laboratory tests for characterization of Histosols*. [In:] M. Stelly (ed.), *Histosols: their characteristics, classification and use*, SSSA Spec. Pub. 6 Madison, WI., 11–20.
- [21] Łachacz, A., Nitkiewicz, M., Kalisz, B., 2009. *Water repellency of post - boggy soils with a various content of organic matter*. *Biologia*, 64, 3: 634–638, DOI: 10.2478/s11756-009-0096-5.
- [22] Marttila, H., Klove, B., 2010. *Dynamics of erosion and suspended sediment transport from drained peatland forests*. *Journal of Hydrology*, 388: 414–425, DOI: 10.1016/j.jhydrol.2010.05.026.
- [23] Matyka-Sarzyńska, D., Sokołowska, Z., 2005. *Physicochemical properties of mucks at different stage of secondary transformation* (in Polish). *Acta Agrophysica, Monographies*, 123: 1–69.
- [24] Migoń, P., Latocha, A., Parzóch, K., Kasprzak, M., Owczarek, P., Witek, M., Pawlik, Ł., 2011. *Contemporary geomorphic system of the Stołowe Mountains* (in Polish). [In:] T. Chodak, C. Kabała, J. Kaszubkiewicz, P. Migoń, J. Wojewoda (eds.), *Geoekologiczne warunki środowiska przyrodniczego Parku Narodowego Gór Stołowych*. WIND, Wrocław, pp. 1–52.

- [25] Minayeva, T., Sirin, A., Bragg, O. (eds.), 2009. *A Quick Scan of Peatlands in Central and Eastern Europe*, Wetlands International, Wageningen, the Netherlands, 132 pp.
- [26] Minkinen, K., Laine, J., 1996. *Effect of forest drainage on the peat bulk density and carbon stores of Finnish mires' julkaisussa*. Proceedings of the International Workshop on "Northern Peatlands in Global Climatic Change", Hyttiala, Finland, 8–12 October 1995, 1: 242–249. Publication of the Academy of Finland.
- [27] Nichols, D.S., Boelter, D.H., 1984. *Fiber size distribution, bulk density, and ash content of peats in Minnesota, Wisconsin, and Michigan*. Soil Science Society American Journal, 48: 1322–1328, DOI: 10.2136/sssaj1984.03615995004800060024x.
- [28] Nieminen, M., Ahti, E., Koivusalo, H., Mattsson, T., Sarkkola, S., Laurén, A., 2010. *Export of suspended solids and dissolved elements from peatland areas after ditch network maintenance in south-central Finland*. *Silva Fennica*, 44, 1: 39–49, DOI: 10.14214/sf.161.
- [29] Okruszko, H., 1993. *Transformation of fen-peat soil under the impact of draining* (in Polish). *Zeszyty Problemowe Postępów Nauk Rolniczych*, 406, 3–73.
- [30] Pakarinen, P., Gorham, E., 1984. *Mineral element composition of Sphagnum fuscum peats collected from Minnesota, Manitoba and Ontario*. [In:] C.H. Fuchsman, S.A. Spigarelli (eds.), Proc. Int. Symposium on Peat Utilization, Bemidji State University, Bemidji, Minnesota; October 10–13, 1984, 417–429.
- [31] Pietiläinen, O.P., Rekolainen, S., 1991. *Dissolved reactive and total phosphorus load from agricultural and forest basins to surface water in Finland*. *Aqua Fennica*, 21: 127–136.
- [32] Prevost, M., Plamondon, A.P., Belleau, P., 1999. *Effects of drainage of a forested peatlands on water quality and quantity*. *Journal of Hydrology*, 214: 130–143, DOI: 10.1016/S0022-1694(98)00281-9.
- [33] Rantonen, H., Paivanen, J., 1999. *Silvicultural condition on tree stand after thinning on drained peatlands*. *Silva Fennica*, 23, 1: 33–50.
- [34] Sallantausta, T., 1988. *Water quality of peatlands and man's influence on it*. [In:] Symposium On the Hydrology of Wetlands in Temperate and Cold Regions, 2, Joensuu, Publication of the Academy of Finland 5, Helsinki 80 pp.
- [35] Sarkkola, S., Hokka, H., Koivusalo, H., Nieminen, M., Ahti, E., Paivanen, J., Laine, J., 2010. *Role of tree stand evapotranspiration in maintaining satisfactory drainage condition in drained peatlands*. *Canadian Journal of Forest Research*, 40: 1485–1496, DOI: 10.1139/X10-084.
- [36] Wang, X., Westbrook, C., Bedard-Haughn, A., 2016. *Effect of mineral horizons on spatial distribution of soil properties and N cycling in a mountain peatland*. *Geoderma*, 273: 73–82, DOI: 10.1016/j.geoderma.2016.03.012.
- [37] Van Rijn, L.C., 1987. *Mathematical modeling of morphological processes in the case of suspended sediment transport*. Ph.D. thesis, Delft University of Technology, the Netherlands.
- [38] Van Reeuwijk, L.P., 2002. *Procedures for Soil Analysis. 6th edition*. – Technical Paper/International Soil Reference and Information Centre, Wageningen, the Netherlands. 4 pp.
- [39] Varry, E.S., 1988. *The hydrology of wetlands and man's influence on it*. [In:] Symposium On the Hydrology of Wetlands in Temperate and Cold Regions, 2, Joensuu, Publication of the Academy of Finland 5, Helsinki, pp. 41–61.